



## 42 INTRODUCTION

43 In 2009, for the first time since World War II, the U.S. vehicle fleet diminished in size, as 14  
44 million vehicles were scrapped and 10 million new vehicles were sold (Brown 2010). Alongside  
45 a U.S. trend toward lower private-vehicle ownership (Cohen 2012) and a growing popularity of  
46 the shared-use economy (Botsman and Rogers 2010), carsharing is emerging as an alternative  
47 travel “mode” that is both more flexible than transit and less expensive than traditional  
48 ownership. Both peer-to-peer carsharing (through organizations like Getaround and Relayrides)  
49 and business-to-consumer carsharing (through operations like Car2Go and Zipcar) are gaining  
50 ground in urban areas. Worldwide, carsharing organizations operate in over 1,100 cities across at  
51 least 27 countries (Shaheen and Cohen 2013).

52 In North America alone, carsharing systems exist in more than 20 metropolitan areas (Martin and  
53 Shaheen 2011b) and membership levels are over 1 million persons (Shaheen and Cohen 2013).  
54 Shared mobility innovations are rapidly growing, and policymakers may want to incentivize this  
55 special mode far more than we have seen to date. Carsharing generally reduces automobile  
56 dependence and lowers emissions while benefitting its users via lowered travel costs (Litman  
57 2000). For decision makers to better appreciate carsharing’s contributions, it is useful to quantify  
58 the life-cycle implications of a shift toward shared-car use. A life-cycle inventory (LCI) analysis  
59 quantifies the complete energy and atmospheric emissions for the lifetime effects of a product,  
60 process, or activity (USEPA 1995), allowing decision makers to compare alternative activities  
61 and products via common metrics.

62 This paper quantifies life-cycle energy and greenhouse gas (GHG) emissions for the carsharing  
63 mode as compared to one equivalent person-kilometer traveled (PKT) under the traditional  
64 (private vehicle ownership) approach. The calculations recognize the vehicle replacement rate  
65 changes that come with sharing, as well as the fuel efficiency improvements from faster fleet  
66 turnover, travel distances changes, reduced parking demands, and shifts in the use of alternative  
67 modes.

## 68 PRIOR RESEARCH

69 Existing studies have examined the environmental impact of carsharing operations to various  
70 extents, but few consider life-cycle impacts, which include upstream infrastructure and supply  
71 chains (like vehicle manufacture and fuel production). Those that examine life-cycle impacts of  
72 carsharing operations exclude key behavioral impacts, such as members’ reduced travel  
73 distances, as discussed in this section.

74 Martin and Shaheen (2011a) estimated GHG reductions at the household level via an analysis of  
75 11 carsharing organizations in North America and found that, while some carsharing members  
76 increase and others decrease their annual emissions, the net impact is an estimated annual  
77 reduction of -0.58 tons of GHG emissions (CO<sub>2</sub>-equivalent, per member household, per year)  
78 due to observed changes in household driving for North American member households and -0.84  
79 tons of GHG emissions in full impacts per member household per year (including foregone  
80 vehicle purchases). This reduction roughly translates to 11% to 16% of the average American  
81 household’s transport-related GHG emissions per year (USDOT 2009). Using stated preference  
82 survey data from Bremen, Germany and Brussels, Belgium, Ryden and Morin (2005) estimated

83 emissions savings per new member to be 54% in the former and 39% in the latter, based on  
84 lower vehicle travel distances (vehicle-kilometers traveled, or VKT), increased fleet fuel  
85 economies, and increases in public transit use.

86 Martin and Shaheen's (2011a) and Ryden and Morin (2005)'s emissions reductions estimates did  
87 not reflect any land use impacts of carsharing. Using stated preference data from Car2go  
88 members in Ulm, Germany, Finkhorn and Muller (2011) examined both mobile emissions  
89 impacts and changes in parking and roadway infrastructure requirements. They estimated a  
90 reduction of 146 to 312 kg of CO<sub>2</sub> per member per year, taking into account reduced vehicle  
91 ownership and VKT. Switzerland's Mobility Carsharing operation has developed an  
92 environmental inventory tool to assess their fleet's consequences - from vehicle manufacturing  
93 and maintenance, to road use, infrastructure provision, and land use effects. As compared to the  
94 average Swiss passenger car, they estimate that the Mobility Carsharing fleet reduces overall  
95 environmental burdens (including exhaust emissions, fuel consumption, material use for car and  
96 road infrastructure, health damages from road noise, and motor vehicle accidents) by 39% per  
97 vehicle-kilometer travelled (VKT), on a vehicle-to-vehicle comparison, thus ignoring additional  
98 savings from members' reduced VKT (Doka and Ziegler 2001). For modern cars with low  
99 emissions, carsharing did not provide significant reductions of NO<sub>x</sub>, HC, CO, and PM10 (as  
100 compared to clear benefits in CO<sub>2</sub>, noise, accidents, and fuel production. The authors noted that  
101 as vehicles become more fuel efficient, land use aspects (e.g., transportation infrastructure  
102 requirements) become a more significant share of the total environmental burden reduction.

103 Briceno et al. (2004) have extended the scope of life-cycle analysis (LCA) for shared-vehicle  
104 systems by anticipating rebound (in consumption) effects, via the use of input-output analysis (to  
105 derive emissions from added non-transport consumption that comes from the average member's  
106 travel cost savings). They found that if car-sharers in Norway spread their transportation savings  
107 uniformly across non-transport items, the overall rebound effects are small. However, if the  
108 travel-cost savings were spent on air travel, the added (rebound) GHG emissions are high,  
109 demonstrating how moves towards ostensibly more sustainable consumption patterns can have  
110 rather unintended consequences. As Hertwich (2005) notes, carsharing typically reduces local  
111 travel expenditures, but use of those savings in other expenditure categories can have negative  
112 environmental impacts.

## 113 **CARSHARING'S IMPACTS ON ENERGY USE AND GHG EMISSIONS**

114 Life-cycle analysis offers a systematic approach to evaluating the environmental consequences  
115 of carsharing, painting a complete picture of this emerging mode's environmental impacts - as  
116 measured in an equivalent PKT. This "cradle-to-grave" process recognizes resource extraction to  
117 produce the vehicles and fuels, and resource depletion through the vehicle use and disposal  
118 phases. Environmental impacts are numerous along the way: First, vehicle "ownership" (in terms  
119 of vehicles per person) generally falls with carsharing membership, offering environmental  
120 benefits from vehicle production and parking infrastructure savings. Second, carsharing has  
121 impacts on VKT and vehicle utilization rates (and thereby fleet replacement rates), which tends  
122 to reduce fuel consumption (as well as, arguably, road infrastructure needs, though this potential  
123 savings is generally not assessed). Lastly, carsharing shifts many trips previously carried out by  
124 private automobile to transit and non-motorized modes (as well as some trips previously carried  
125 out by non-auto modes to shared cars). As pointed out above, in this paper's literature review,

126 prior studies have examined the environmental impact of carsharing to different extents, but no  
127 study has examined the overall impact of all these behavioral changes associated with carsharing  
128 concurrently (ownership impacts on vehicle production and transportation infrastructure, vehicle  
129 utilization and fleet replacement, and modal shift). This study applies an LCA framework to  
130 comprehensively examine the combined effects on energy use and GHG emissions accounting  
131 for all of these potential traveler behavior shifts.

### 132 **Candidate Households for Carsharing**

133 However, carsharing is not a reasonable option for every traveler. Carsharing membership is  
134 more appealing for those who travel fewer kilometers and reside in higher-density  
135 neighborhoods with good walking, cycling, and transit options (Litman 2000). Thus, carsharing  
136 programs tend to concentrate in metropolitan cores, well served by other modes, where travelers  
137 can and do rely less on private car use than the average traveler (Stillwater et al. 2009). In an  
138 analysis of 13 U.S. regions with carsharing programs, Celsor and Millard-Ball (2007) found that  
139 carsharing neighborhoods are more likely to have higher shares of one-person households and  
140 residents with Bachelor's degrees, more workers commuting by transit and non-motorized  
141 modes, lower vehicle ownership levels, higher density, and more walkable environments than  
142 non-carsharing neighborhoods. Furthermore, carsharing trips are more like to be used for  
143 shopping, personal business, and recreation trips versus commute trips (see, e.g. Millard-Ball et  
144 al. 2005 and Cervero et al. 2007), and members' average trip distances are shorter than those of  
145 non-members (Cervero et al. 2007).

146 Thus, while carsharing is not an omnipresent and universally feasible travel option, it does  
147 appeal to various populations. Frost and Sullivan (2010) estimated that car owners who drive  
148 12,000 miles (7,460 km) per year at an average speed of 30 mi/hr can save \$1,834 by switching  
149 to a carsharing service (with those driving less than 12,000 miles reaping even greater savings).  
150 Looking specifically at the San Francisco Bay Area, Duncan (2011) estimates that as much as  
151 one-third of those households have vehicle usage patterns that would save money via carsharing.  
152 Others are not as optimistic: Schuster et al. (2005) estimate that in Baltimore, Maryland, 4.2% to  
153 14.8% of vehicles would be less expensive to share than to. If estimates from the Bay Area and  
154 Baltimore are applied to urban areas throughout the US (taking into account that 80% of the US  
155 population now resides in urban areas (Census 2010), the range of potential carsharing members  
156 nationwide covers a wide spectrum: from 3% up to 26% of persons. Surveying 26 existing  
157 organizations in North America, Shaheen et al. (2006) estimate that market potential for  
158 carsharing is 10% of adults 21 and older. In addition to existing neighborhood infrastructure and  
159 household demographics, policy can also affect the adoption of carsharing. Using stated  
160 preference survey from Palermo, Italy, Catalano et al. (2008) developed a multinomial logit  
161 (MNL) model which simulated that carsharing activity could increase up to 10% with policies  
162 which increase parking fees, add reserved parking areas for carsharing and carpooling users,  
163 and closing off specific traffic zones for high-emissions vehicles.

164 For members who actively participate in carsharing, the adoption of carsharing behavior has  
165 quantifiable effects on vehicle ownership rates, VKT, and modal shift to and from transit and  
166 non-motorized modes. The energy and GHG impacts of these vehicle ownership and travel  
167 behavior shifts are discussed in detail below.

168 **Vehicle Ownership Impacts**

169 Within carsharing households, early studies estimate that vehicle ownership can be reduced by  
170 about 40% to 44% (Whitelegg and Britton 1999, Meijkamp 1998). Zhou and Kockelman (2011)  
171 surveyed Austin, Texas households in 2008 and found that 21% of those surveyed (following  
172 population correction) would expect to give up/release at least one of their private held vehicles  
173 upon joining a carsharing organization. A 2008 nationwide survey found that after carsharing,  
174 US households reduced their overall vehicle ownership by 49%, with most of this shift from one-  
175 car households to no-car households (Martin and Shaheen 2011b). In the San Francisco Bay  
176 Area, Cervero et al. (2007) looked at the longer term effects of membership in City Carshare and  
177 found that vehicle shedding effects level off with length of membership. A survey 4 years  
178 following the program's establishment found that the net vehicle shedding effects (as compared  
179 to non-member households) is about 10 vehicles per 100 households. Martin et al. (2010) also  
180 concluded that the combined effect of vehicles shed and vehicles avoided translates to each  
181 carsharing vehicle serving in lieu of 9 to 13 privately owned vehicles. A first-year look in  
182 Philadelphia estimates that each PhillyCarShare vehicle replaced, on average, 23 private vehicles  
183 (Lane 2005). Other North American studies have estimated the vehicle replacement rate closer to  
184 one carsharing vehicle per 15 privately owned vehicles (Millard-Ball et al. 2005, Econsult 2010,  
185 Frost & Sullivan 2010, Stasko et al. 2013).

186 **Impacts on Vehicle-Kilometers Traveled (VKT)**

187 Upon joining a carsharing operation, households typically travel by car less than prior to joining  
188 carsharing. When use of a vehicle involves reserving a vehicle in advance and the costs of  
189 operating a vehicle are made more apparent (generally with a by the minute charge in most car-  
190 share operations), households tend to decrease their use of vehicles. Comparing similar  
191 households in Montreal, Sioui et al. (2012) found that households who subscribe to and active  
192 use a carsharing organization utilize a car 3.7 times less than neighbors who do not subscribe to  
193 these services.

194 However, estimates of how much households reduce their auto travel distances vary greatly.  
195 Sperling et al. (2000) estimate carsharing reduces VKT by 30-60%. Frost and Sullivan (2010)  
196 estimate carsharing members drive 31% fewer kilometers upon joining a carsharing service.  
197 Cervero et al. (2007) looked at members of City CarShare in San Francisco and found that in the  
198 long term, carsharing members reduced their annual VKT by 67%. Martin and Shaheen (2011b)  
199 found through a North American survey that the average VKT by respondents decreased 27%  
200 after joining carsharing (from 6468 km/year to 4729 km/year). In Europe, these impacts seem to  
201 be even greater as Muheim (1998) estimates that members of Mobility Carsharing Switzerland  
202 drove 72% fewer kilometers after their first year of joining the program and Meijkamp (1998)  
203 reports that members of carsharing organizations in The Netherlands drove 33% fewer miles  
204 after becoming car-sharers. Ryden and Morin (2005) used stated preference surveys and found  
205 that, on average, carsharing members in Bremen, Germany and Brussels, Belgium reduced their  
206 VKT by 45 and 28%, respectively.

207 **Impacts on Energy Consumption During Use Phase**

208 In addition to reducing use phase energy demand by reducing VKT, members of car-share  
209 operations also tend to drive more fuel efficient vehicles than non-car-share members. Meijkamp  
210 (1998) estimate that shared cars are approximately 24% more fuel efficient than the average car  
211 in the Netherlands. Martin and Shaheen (2011a) also found that carsharing vehicles are more fuel  
212 efficient than the vehicles they replaced, with the carsharing fleet averaging 13.9 km per liter  
213 (32.8 mpg) and the vehicles they replaced averaging 9.8 km per liter (23.3 mpg). Using stated  
214 preference data from Germany and Belgium, Ryden and Morin (2005) estimated that the average  
215 carsharing vehicle is 17% more fuel efficient than the average privately owned vehicle. This  
216 phenomenon can probably be attributed to the faster replacement rate of car-share vehicles since  
217 they have higher utilization rates. The average privately owned new vehicle in the U.S. is owned  
218 for 71.4 months (or approximately 6 years) before being “replaced”, which may be via sale as a  
219 used vehicle, trade-in (when acquiring a newer or different vehicle), shedding an unneeded  
220 vehicle, or a serious crash (Seng 2012). On the other hand, due to more VKT and faster wear and  
221 tear, the commercial car-share operations replace cars every 2 to 3 years (Mont 2004). With  
222 government mandates like CAFE standards and increasing fuel prices, newer vehicles, on  
223 average, are more fuel efficient (and smaller) than older fleets, contributing to a more fuel  
224 efficient shared fleet compared to a privately owned fleet.

## 225 **Impacts on Parking Infrastructure Demand**

226 Reduced car ownership has potential impacts on infrastructure requirements, particularly  
227 parking. Most governing authorities’ interest in promoting carsharing is motivated by parking  
228 demand reduction (Millard-Ball et al. 2005). While numerous studies qualitatively link reduced  
229 vehicle ownership and parking demand (see, e.g., Millard-Ball et al. [2005] and Martin et al.  
230 [2010]), few studies have quantified the magnitude of that impact. The 1-to-15 shared-vehicle-to-  
231 private-vehicle replacement rate discussed earlier does not directly translate to a parking impact  
232 in high-demand areas, since many car-share participants use transit and other non-auto modes for  
233 commute trips (Celsor and Millard Ball 2007), and hence much of the parking reduction would  
234 occur in private garages and parking lots. A 2004 study in the U.K. surveyed employers and  
235 found that spaces fell from 0.79 spaces per staff member to 0.42 spaces per staff member after  
236 starting a carsharing program (Department for Transport 2004). Looking at carsharing and  
237 parking at the building scale in Toronto, Engel-Yan and Passmore (2013) found that buildings  
238 with dedicated carshare vehicles required 50% fewer parking spaces than those without such  
239 dedications. Using survey data from Ithaca Carshare, Stasko et al. (2013) estimated that program  
240 participants’ on-street parking needs or demands fall by 26 to 30%, depending on day of week  
241 and time of the day.

## 242 **Impacts on Other Modes of Transportation**

243 So how do car-share members pursue trips while reducing vehicle ownership and cutting VKT?  
244 Overwhelming, studies point to increase use in non-motorized modes and transit. In the  
245 Netherlands, Meijkamp (1998) reports 14% increase in bicycling, 36% increase in rail transit  
246 use, and 34% increase in bus transit use among carsharing members. In Germany and Belgium,  
247 Ryden and Morin (2005) estimate that carsharing members use public transportation 35 to 47%  
248 more during weekdays. In Montreal, Canada, households who subscribe to carsharing services  
249 use public transportation 55% more often than neighbors who own one private vehicle (Sioui et  
250 al. 2012). In the US, a second year evaluation of CarSharing Portland found members reporting

251 25% increase in walking, 10% increase in bicycling, and a 14% increase in public transit use  
 252 (Cooper et al. 2000). Similar results can be seen in Philadelphia after one year of joining Philly  
 253 CarShare, 19% of members reported more walking, 8% reported more cycling, and 18% reported  
 254 more transit use (Lane 2005). In a survey of 13 car sharing operations in North America, Martin  
 255 and Shaheen (2011c) found the impact on transit use was statistically insignificant after joining  
 256 car sharing programs but net use of walking, biking, and carpooling modes increased 2%, 7%,  
 257 and 3%, respectively.

258 **ANALYSIS AND RESULTS**

259 The total impact of carsharing on energy use and GHG emissions as compared to an equivalent  
 260 PKT in a private automobile (the functional unit in this study) is the combined effect from all of  
 261 these different dimensions of travel behavior, vehicle technology, and infrastructure change. This  
 262 analysis presents three different scenarios to examine the sensitivity of reduction in total life-  
 263 cycle energy and GHG emissions for a candidate household member (one who travels shorter  
 264 total distances and resides in higher-density urban neighborhoods, with good walking, cycling,  
 265 and transit services) upon joining a carsharing organization. Table 1’s results for low-impact  
 266 (pessimistic), medium-impact (likely), and high-impact (optimistic) scenarios are based on  
 267 multiple input factors (as shown in Table 1’s first column). The values and ranges of these inputs  
 268 come from the studies discussed earlier, in this paper’s Impacts section, where the low-impact  
 269 scenario represents results from the most pessimistic estimate from prior studies, the high-impact  
 270 scenario represents the most optimistic estimate, and the medium-impact scenario reflects a  
 271 commonly agreed-upon estimate from multiple previous studies (in the better-studied impact  
 272 categories) or a single study’s value lying between the low- and high-impact estimates.

273 **Table 1. Effect of Carsharing on Travel Behavior, Infrastructure Demand, and Other**  
 274 **Modes**

<b>Input</b>	<b>Low Impact</b>	<b>Source(s)</b>	<b>Med Impact</b>	<b>Source(s)</b>	<b>High Impact</b>	<b>Source(s)</b>
Carsharing Market Potential (% of US adult population)	3.0%	Schuster et al. (2005)	10.0%	Shaheen et al. (2006)	26.0%	Duncan (2011)
% Reduction in Private Vehicles Owned	10.0%	Cervero et al. (2007)	21.0%	Zhou and Kockelman (2011)	49.0%	Martin and Shaheen (2011b)
Private Vehicle Replacement Rate for Each Car-Sharing Vehicle	9	Martin et al. (2010)	15	Millard-Ball et al. (2005), Econsult (2010), Frost and Sullivan (2010), Stasko et al. (2013)	23	Lane (2005)
% Reduction in VKT	27.0%	Martin and Shaheen (2011b)	31.0%	Frost and Sullivan (2010)	67.0%	Cervero et al. (2007)
% Fuel	17.0%	Ryden and	24.0%	Meijkamp	43.5%	Martin and

Efficiency Improvement		Morin (2005)		(1998)		Shaheen (2011a)
% Reduction in Public Parking Demand	26.0%	Stasko et al. (2013)	38.0%	Department for Transport (2004)	50.0%	Engel-Yan and Passmore (2013)
% Increase in Rail Transit Use	0.0%	Martin and Shaheen (2011c)	8.0%	Cooper et al. (2000), Lane (2005)	36.0%	Meijkamp (1998)
% Increase in Bus Transit Use	0.0%	Martin and Shaheen (2011c)	8.0%	Cooper et al. (2000), Lane (2005)	34.0%	Meijkamp (1998)
% Increase in Bicycling	7.0%	Martin and Shaheen (2011c)	9.0%	Cooper et al. (2000), Lane (2005)	14.0%	Meijkamp (1998)
% Increase in Walking	2.0%	Martin and Shaheen (2011c)	19.0%	Lane (2005)	25.0%	Cooper et al. (2000)

275

276 The energy use and GHG emissions impacts are estimated relative to the base case (“Before”  
 277 scenario) of private vehicle ownership (prior to joining a car-share organization). As discussed  
 278 previously, potential carsharing participants exhibit different travel behaviors than the average  
 279 motorist. The calculations on energy and emissions impacts as a result of mode shift are based on  
 280 initial mode shares of “likely” candidates for carsharing membership, based on findings in Celsor  
 281 and Millard-Ball (2007) and Cervero et al. (2007).

282

**Table 2. Base Mode Split for Candidate Carsharing Members**

	<b>Mode Split</b>
Private Car	33.6%
Rail Transit	19.4%
Bus Transit	11.6%
Bike	3.8%
Walk	31.6%

283 The impacts of vehicle operation changes are estimated as a result of reduction in VKT in Table  
 284 1 and are based on per PKT estimates of a conventional sedan (2005 Toyota Camry with a  
 285 combined fuel economy of 25 mpg) from Chester and Horvath (2009). As discussed in the  
 286 Impacts on Energy Consumption During Use Phase section, the average private vehicle replaced  
 287 by a carsharing vehicle averages 23.3 mpg (Martin and Shaheen 2011a), making this estimate  
 288 slightly conservative. The impacts of vehicle manufacturing and maintenance changes are a  
 289 result of the private vehicle replacement rate in Table 1 and are based on per PKT estimates of a  
 290 2005 Toyota Camry from Chester and Horvath (2009). The impacts of parking infrastructure  
 291 demand decrease is a result of percent reduction in public parking demand in Table 1 and are  
 292 based on the per PKT estimates of a total inventory of 820 million parking spaces in the US  
 293 including for-pay parking spaces, commercial spaces, and on-street parking from Chester et al.  
 294 (2010). The impacts of decreased fuel production are a result of the percent fuel efficiency  
 295 improvement in Table 1 and are based on per PKT estimates for a 2005 Toyota Camry in Chester  
 296 and Horvath (2009). The energy and GHG emissions impacts from increased rail transit use are



297 based on an average of San Francisco Muni operations in the Bay Area and Green Line  
 298 operations in Boston (since carsharing members living in the city core are more likely to use  
 299 light rail over heavy commuter rail). Those for bus transit use are based on operations of a  
 300 typical 40 ft diesel bus (with combined fuel economy of 4.3 mpg) during peak congestion hours  
 301 as reported in Chester and Horvath (2009). Lastly, the impacts of increased use of walk and bike  
 302 (assumed non-electric) modes are from per PKT estimates in Dave (2010).

303 **Table 3. Energy and GHG Emissions per Equivalent Private Vehicle PKT**

	Energy (MJ)					GHG (g CO <sub>2</sub> equiv)				
	Per PKT	Before	After-Low	After-Med	After-High	Per PKT	Before	After-Low	After-Med	After-High
Vehicle Operation	2.1	2.1	1.53	1.06	0.35	144.15	144.15	105.23	72.61	23.96
Vehicle Manufact. & Maintenance	0.37	0.37	0.04	2.74E-03	1.19E-04	29.76	29.76	3.31	0.22	0.01
Parking Infrastructure	0.5	0.5	0.37	0.23	0.11	46.6	46.6	34.48	21.38	10.69
Fuel Production	0.24	0.24	0.20	0.15	0.09	24.18	24.18	20.07	15.25	8.62
Increased Rail Transit Use	1.61	0	0.00	0.07	0.33	122.33	0	0.00	5.65	25.44
Increased Bus Transit Use	0.67	0	0.00	0.02	0.08	51.56	0	0.00	1.43	6.08
Increased Bike Mode Use	0.2	0	0.00	0.00	0.00	20.63	0	0.16	0.21	0.33
Increased Walk Mode Use	0.06	0	0.00	0.01	0.01	20.63	0	0.39	3.68	4.84
<b>Totals</b>		<i>3.21</i>	<i>2.15</i>	<i>1.55</i>	<i>0.98</i>		<i>244.69</i>	<i>163.64</i>	<i>120.44</i>	<i>79.96</i>
<b>Total % Reduction</b>			<i>33.2%</i>	<i>51.8%</i>	<i>69.5%</i>			<i>33.1%</i>	<i>50.8%</i>	<i>67.3%</i>

304

305 As seen in Table 3, for a traveler who drives relatively few miles each year and lives in a denser  
 306 urban neighborhood with good access to transit and non-motorized modes, joining a carsharing  
 307 organization can reduce his/her energy use and GHG emissions 33 to 70%. In the most likely  
 308 scenario, both inventories are reduced about 51% after a candidate traveler joins a carsharing  
 309 organization. It is apparent that the energy use and GHG reductions are dominated by changes in  
 310 vehicle operations, which is a result of reduced trips and travel distances in an automobile. In  
 311 other words, the most important contributor to carsharing's lowered impacts is avoided travel  
 312 and travel shifted to non-auto modes. While carsharing can increase the service of underutilized  
 313 vehicles (with more vehicles replaced due to miles driven, rather than age-related factors - like  
 314 rust or outdated design), the primary driver behind environmental benefits seem to arise out of a  
 315 traveler's need to plan for travel and awareness of the cost of automobile travel, since most  
 316 carsharing services require reservations and operate on a pay-by-the-minute basis.

317 Following vehicle operations, the biggest energy and GHG emissions reductions can be seen in  
318 parking infrastructure demand, followed by fuel use decreases, which come from reduced auto  
319 ownership, shifted modes, and vehicle technology improvements. Even if one considers only  
320 impacts to public parking infrastructure, carsharing's life-cycle energy and emissions savings are  
321 substantial. Despite the literature's emphasis on vehicle ownership reduction and vehicle  
322 replacement ratios, vehicle manufacture and maintenance have a relatively small impact on total  
323 energy use and GHG emissions per equivalent private vehicle PKT.

324 The biggest inventory changes from trips shifting to non-automobile modes emerge from transit  
325 use changes, particularly to the rail mode. In the most likely (medium-impact) scenario, the  
326 estimated increase in energy and GHG emissions from increased use of all other modes (rail,  
327 bus, bike, and walk) is less than savings from lowered parking demands. However, transit  
328 impacts are quite sensitive to occupancy assumptions. As noted earlier, environmental impact  
329 estimates from increased rail use are based on San Francisco's Muni (light-rail) and Boston's  
330 Green Line (light rail) operations, and bus use impacts are based on peak-hour diesel bus  
331 operations. When using the worst-case, low-occupancy assumptions (5 passengers per bus and  
332 25% occupied seats on light rail) from Chester and Horvath (2009), the rise in energy use as a  
333 result of increased transit use is estimated to be 0.25 MJ per equivalent private-vehicle PKT and  
334 the rise in GHG emissions is estimated to be 20.6 gm CO<sub>2e</sub> per equivalent private-vehicle PKT.  
335 In other words, when transit occupancies are assumed to be low, the corresponding increase in  
336 environmental impacts as a result of increased transit trips is comparable to the decrease in  
337 environmental impacts from reduced public parking needs, as a result of a candidate household  
338 joining a carsharing program, on an equivalent PKT basis.

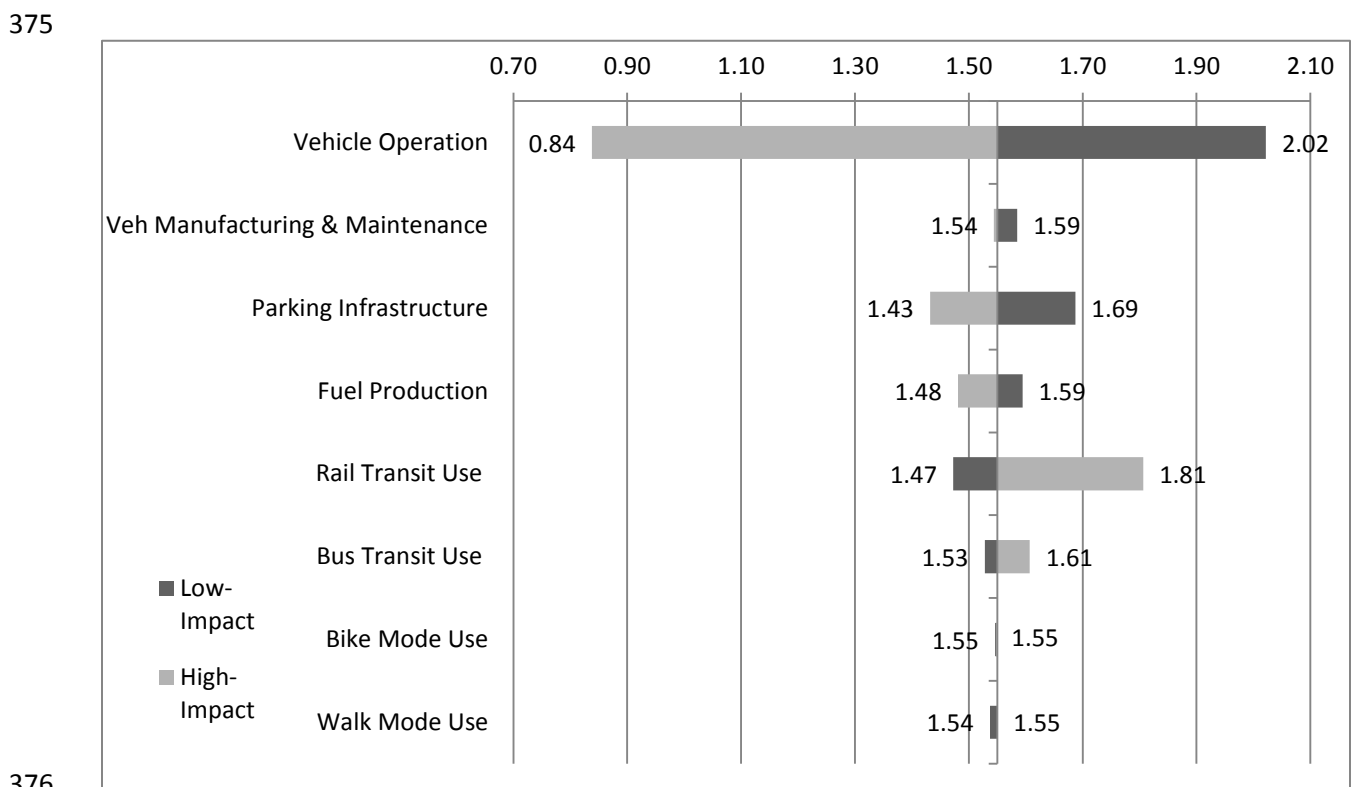
339 Potential reductions in energy use and GHG emissions across all US households as a result of all  
340 candidate households joining carsharing organizations are in the range of 1 to 18 percent, based  
341 on Table 1's nationwide carsharing market potential numbers, with the most likely scenario  
342 showing an approximate net 5 percent reduction in energy use and GHG emissions in local  
343 household transportation if all candidate households join carsharing organizations (as compared  
344 to those households using private, non-shared vehicles). While this analysis assumes that 3.0 to  
345 26.0 percent of US households could be candidate members for carsharing organizations, as of  
346 2013, presently there are just about 800,000 carsharing members in the United States (Steinberg  
347 and Vlastic 2013), or less than half of one percent of the nation's 210 million licensed drivers,  
348 and their 246 million registered (non-commercial) vehicles (USDOT 2011).

349 It is important to note that while these calculations include direct rebound effects as a  
350 consequence of joining a carsharing organization, in the form of increased transit and  
351 nonmotorized trips, they do not account for indirect, economy-wide rebound effects of the  
352 avoided and shifted mode trips. From a household perspective, transportation expenditures  
353 savings will likely be used to purchase other products and services, which also require energy  
354 and have environmental impacts. Since indirect rebound effects are difficult to calculate (as a  
355 result of a whole host of second-order effects), estimated impacts from energy and GHG  
356 emissions indirect rebound vary widely. Experts estimate these effects to be as little as 5 to 15%  
357 (Thomas and Azevedo 2013, Druckman et al. 2011) to as much as 35 to 40% (Sorrell 2007).  
358 Thus, with indirect rebound effects considered, the likely total life-cycle inventory energy and  
359 GHG emissions savings from all U.S. candidate households joining carsharing organizations is

360 arguably in the range of 3 to 5% of all local household transport-related energy use and  
 361 emissions.

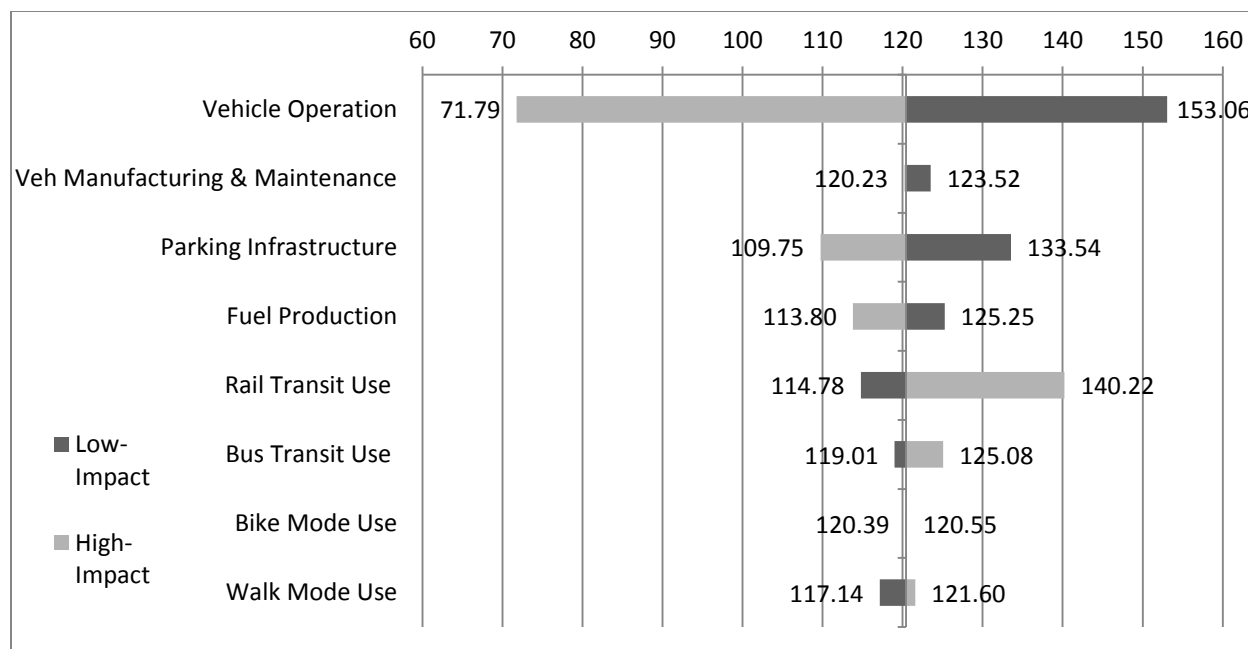
362 Of course, each component of travel behavior change, infrastructure demand change, and  
 363 technology change impacts the total reduction differently, as seen in the tornado graphs shown  
 364 below. In Figures 1 and 2, the baseline (y-axis) value represents the most likely energy use (1.55  
 365 MJ) and GHG emissions (120.44 g CO<sub>2e</sub>) per equivalent private-vehicle PKT, assuming medium-  
 366 level impacts (from Table 1) across all impact categories. The bars associated with each impact  
 367 category show the range of energy use and GHG emissions associated with changing that impact  
 368 from low to high, while all other impact categories remain at medium levels. These graphs  
 369 illustrate the sensitivity of results to the estimates in each impact category. As seen in Figures 1  
 370 and 2, results are most sensitive to carsharing's assumed impacts on VKT, followed by assumed  
 371 increases in rail transit use and decreased demand for public parking. It is interesting to note that,  
 372 while the first two topics are well covered in past literature, the number of studies examining  
 373 carsharing's parking infrastructure impacts is limited.

374 **Figure 1. Impact of Inputs on Energy Use (MJ) per Equivalent Private Vehicle PKT**



377

378 **Figure 2. Impact of Inputs on GHG Emissions (g CO<sub>2</sub> equiv) per Equivalent Private**  
 379 **Vehicle PKT**



380

### 381 CONCLUSIONS AND EXTENSIONS

382 The benefits of carsharing have been touted in many previous studies, from reductions in vehicle  
 383 ownership levels to increased transit use. However, few studies have examined the life-cycle  
 384 impacts of carsharing (including upstream supply chains for vehicles and fuel), and they exclude  
 385 infrastructure and/or shifted-mode components. Using estimates from a wealth of previous  
 386 carsharing studies, this study quantifies the life-cycle reductions in energy and GHG emissions  
 387 of carsharing as compared to an equivalent PKT in a private vehicle, combining the effects of  
 388 reduced vehicle ownership, reduced vehicle distance traveled, fleet-level fuel efficiency  
 389 improvements, reduced parking infrastructure demand, and trips shifted to no-auto modes. For a  
 390 traveler that meets the criteria of a good candidate for carsharing, joining a carsharing  
 391 organization is predicted to decrease his/her transportation energy use and GHG emissions by  
 392 51%, with the biggest reduction coming from decreased vehicle operations as a result of avoided  
 393 VKT or mode shifts. A decrease in parking infrastructure demand also contributes to significant  
 394 reductions in energy use and GHG emissions, as recognized through the LCA process employed  
 395 here but neglected in studies that emphasize vehicle operations. Additional energy and GHG  
 396 emissions as a result of carsharing members' increased use of transit and non-motorized modes  
 397 are estimated to be insignificant when compared to the savings from avoided private-vehicle  
 398 VKT, even under worst-case scenarios, with low transit occupancy rates. Across all US  
 399 households, this translates to a total energy use and GHG emissions reduction of approximately  
 400 5% for local household transportation activities and as little as 3%, once rebound effects (from  
 401 expenditure of saved funds on other consumer items) are considered.

402 It is important to note that a comprehensive LCA of the environmental impacts of carsharing  
 403 generally relies on estimates from prior studies, so any biases or limitations in those prior studies  
 404 carry forward to this analysis. These biases include the fact that some prior studies rely on stated  
 405 (rather than revealed) preference data (see, e.g. Zhou and Kockelman 2011, Martin and Shaheen  
 406 2011b, Ryden and Morin 2005), in which respondents may over- or under-estimate actual

407 behavioral shifts of carsharing membership. However, by relying on estimates averaged from a  
408 wide range of past studies and providing low-, medium-, and high-level scenarios, the aggregate  
409 energy and GHG emissions impacts estimated here have sought to minimize any biases of past,  
410 individual studies.

411 In averaging across studies that do not distinguish between station-based and free-floating  
412 carsharing services, this LCA does not differentiate the energy and GHG impacts of these two  
413 service types. Previous studies suggest that station-based carsharing services are mostly used for  
414 shorter trips, with shorter parking durations (Costain et al, 2012, Barth and Shaheen 2002), while  
415 free-floating carsharing services tend to be used for a wider variety of trip purposes and distances  
416 (Schmoller et al., 2014). In terms of mode substitution, free-floating carsharing services are  
417 generally found to be more substitutable for transit, walking, and cycling modes than are station-  
418 based systems (Ciari et al., 2014, Le Vine et al., 2014). Moreover, some potential impacts are not  
419 yet quantified here. For example, carsharing may impact roadway construction, lighting, and  
420 maintenance demands, and associated technologies are evolving. Improved catalytic converters  
421 and electrified shared fleets, possibly reliant on renewable feedstocks for their power, may  
422 improve carsharing's contributions to lowered energy demands and emissions. Finally, it should  
423 be noted that this study compares a shared fleet of conventional (internal combustion engine)  
424 sedans to the average U.S. passenger vehicle's use. With smaller, hybrid and electric vehicles  
425 growing in popularity, carsharing's energy and GHG emissions savings will probably grow.

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